Bacterial and Biodeterioration Analysis of the Waterlogged Wooden Lacquer Plates from the Nanhai No. 1 Shipwreck

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Abstract: To protect the lacquer plates from the Nanhai No. 1 shipwreck from being corroded by microorganisms, a series of studies were conducted on the four water-stored samples. The water samples were collected from the vessel where the lacquerware was stored in June and December 2017. In our study, high-throughput sequencing was conducted to reveal predominant bacterial communities. Then, three different media were used to isolate the dominant bacteria, and the 16S rRNA genes were sequenced. Next, we tested the degradation activity of lignin and cellulose by the isolated bacterial strains. After being cultured on a medium containing carboxymethylcellulose (CMC), almost all the isolated strains (except Microbacterium sp. NK-NH4, Ochrobactrum sp. NK-NH9, and Bacillus megaterium NK-NH10) showed the capacity for cellulose degradation. In addition, the lignin peroxidase (Lip) and laccase activity of the strains were shown by culturing the strains on one medium with azure and on another medium with Remazol brilliant blue. The results indicated that the Lip activity of all the strains was low, whereas the laccase activity of Microbacterium sp. NK-NH4, Bacillus tequilensis NK-NH5, Bacillus subtilis NK-NH6, Bacillus megaterium NK-NH10, and Bacillus velezensis NK-NH11 was relatively high. Finally, we tested the bacteriostatic efficacy of four biocides—Preventol® D7, BIT 20N, P91, and Euxyl® K100. We found that most strains were sensitive to D7 and 20N, while K100 had almost no impact.

Keywords: Nanhai No. 1 shipwreck; lacquerware; high-throughput sequencing; ligninolytic enzymatic activity; cellulolytic enzymatic activity; biocide effectiveness; bacterial diversity

1. Introduction

Nanhai No. 1 was a ship that carried a large amount of treasure for ocean-going trade along the Marine Silk Road [1]. It sank off the coast during the Southern Song Dynasty (1127–1279 AD) and had been resting at the bottom of the sea for decades until 2007. The discovery of the Nanhai
No. 1 shipwreck is an unprecedented achievement and a historical landmark for Chinese underwater archaeology [2].

In recent years, several lacquer plates and many other archaeological woods have been excavated, raised, and conserved. For instance, in 1904, a site at Oseberg was excavated and conserved in Vestfold, Norway, along with a collection of wooden objects. Subsequently, the wood became powder, in the worst cases [3]. In China, a polychrome lacquer plate was excavated from a burial site and was stored in the Museum of Qin Shihuang Terracotta Horses and Warriors in Xi’an, Shanxi province, China. As a result of age, the wooden padding of the polychrome lacquer plate has completely decayed, while the paint film is completely fixed on the soil. Luckily, the shape of the lacquer plate remains intact, and the color of the painting is still bright [4]. Moreover, in March 2006, a lacquer plate was excavated from Tomb No. 10 (YLM10) of Yizheng, an important Early Han cemetery located in Yangzhou, Jiangsu province, China. Although YLM10 was badly damaged by the local brick factory, only one piece of the lacquered wooden plate was in good condition [5]. However, more than one hundred lacquer plates excavated from Tomb No. 1 of Fengpeng Ridge cemetery, located in Changsha, Hunan province, China, are in danger. Most of these plates were seriously distorted and badly mutilated when they were discovered in March 2006 [6].

Archaeological wood discovered in archaeological sites is a valuable archaeological find and tells us a lot about history; however, archaeological wood findings are rare because of long-term, microbiological biodegradation and unavoidable interactions with the environment [7]. Luckily, many scientists and archaeologists have been paying close attention to protecting archaeological wood.

Most of the objects from the Oseberg ship were treated with alum (KAl(SO₄)₂·12H₂O) to maintain the structural integrity of the wood. However, the release of sulfuric acid has reduced these artifacts to their current, mechanically weakened state, and, at present, the wood is highly acidic (pH ≤ 2). Previous analyses have revealed that the carbohydrate component of the wood is almost completely absent [3]. What is the condition of wood found at other archaeological sites? In the Mausoleum of the First Qin Emperor, where the polychrome lacquer plate was excavated, the destruction of archaeological wood has been extremely rapid because environmental factors, such as ambient humidity, have changed a lot since the excavation. The lacquer plates along with the soil around them were covered with a film to retain the moisture, and PEG200 and PU emulsion were also used [4]. However, conditions were different for the lacquer plates excavated from Tomb No. 1 of Fengpeng Ridge cemetery. Due to its low moisture content and high mechanical strength, one of the lacquer plates was dried to keep it in its initial shape [6].

In addition to lacquer plates, many other wooden relics, such as bamboo slips, seem to face similar preservation challenges. The Jiandu Museum in Changsha, Hunan, houses a collection of bamboo slips belonging to the Three Kingdoms period—these slips were excavated in the Marco Pagoda. Due to the bamboo slips being buried underground for a long time before being unearthed, they were generally in a saturated state. A series of experiments have been conducted, and the results show that microorganisms contributed to much of the degradation of the bamboo slips [8]. In addition, another waterlogged archaeological wood called Xiaobaijiao No. 1 was salvaged from the seabed. The whole wooden ship suffered from biodegradation [9].

Generally, the protection process of waterlogged archaeological wood includes washing and desalting, sterilization, swelling, drying, padding, and painting [6,10]. Therefore, in our study, we are at the very first step of the whole process, and the lacquerware is stored in distilled water. Bacteria are still actively biodegrading the waterlogged archaeological wood, causing changes in the physical, chemical, and mechanical properties of the wood along with other environmental conditions (temperature, humidity, etc.) [11]. The environments suitable for wood preservation are waterlogged, extremely dry, or freezing, because microbial activity is drastically reduced under these conditions [7]. In this study, we aim to analyze the microbial degradation of the lacquerware under the condition of water saturation and hope to present some helpful protection measures.
2. Materials and Methods

2.1. Water Sample Collection

The lacquerware from the Nanhai No. 1 shipwreck was kept in a refrigerator at 4 °C. Some portions were stored in distilled water, and the remainder were covered by wet gauze.

Four water samples storing the lacquerware were used for the subsequent experiments. Sample No. 9 was collected from the distilled water storing the lacquer plate, named 2016NHIT0501@935. Samples No. 10 and No. NHWS03 were from the water storing the lacquerware, named 2015NHIT0202©:58. Sample No. NHWS04 was from the water storing the lacquerware, named 15C10©:0068. Samples No. 9 and No. 10 were collected in June 2017. Samples No. NHWS03 and No. NHWS04 were collected in December 2017. All the water samples were taken with 50 mL sterile centrifuge tubes held in ice baths and brought to the laboratory for analysis. (Shown in Figure S1)

2.2. High-Throughput Sequencing

DNA extraction of the collected water samples (No. 9, No. 10, No. NHWS03, and No. NHWS04) was performed using the PowerWater DNA Isolation kit (MO BIO Laboratories, Carlsbad, CA, United States), following the manufacturer’s protocol. The total DNA was extracted from 300 mL of each water sample and dissolved in 100 µL of the elution solution. After DNA extraction, the total DNA was sent to the Beijing Novogene Genome Sequencing Company and amplified by amplicon. As a result, the bacterial community contained in the water sample was analyzed. The raw sequencing data generated from this study were deposited into the National Center for Biotechnology Information (NCBI) Sequence Read Archive (SRA) under the accession numbers SAMN10147289–SAMN10147296 and SAMN10147297–SAMN10147304 for the bacterial reads in June and December 2017, respectively.

2.3. Isolation of Dominant Bacteria

Three different types of media were used to isolate the dominant bacteria from samples No. NHWS03 and No. NHWS04: (i) Luria Broth (LB) agar medium, (ii) MacConkey agar (MAC) medium, 20 g peptone, 5 g bile salt, 5 g NaCl, 17 g agar, 10 g lactose, 10 mL crystal violet in aqueous solution (0.01%), 5 mL neutral red in aqueous solution (0.5%), and 1 L distilled water, and (iii) M2 agar medium, 5 g CH₃COONa, 0.5 g peptone, 0.5 g yeast extract, 0.5 g glucose, 0.5 g sucrose, 0.5 g soluble starch, 0.05 g trisodium citrate, 0.05 g malic acid, 0.05 g potassium sodium tartrate, 1 g NH₄NO₃, 0.2 g NH₄Cl, 15 g agar, and 1 L distilled water (pH = 7.6) [12]. Each sample was spread onto these three different plates and incubated at 28 °C with daily observation. The colonies with different morphologies and appearances were transferred to fresh plates to obtain pure isolates for DNA extraction and to study their physiological and chemical characteristics.

2.4. DNA Extraction and Sequencing of the 16S rRNA Gene

The DNA extraction of pure strains isolated from the water samples was performed by the hexadecyltrimethylammonium bromide (CTAB) method [13]. The bacterial 16S rRNA genes were amplified using the 341f/907R primers. The primer sequences were 341f (CCTACGGGAGGCAGCAG) and 907R (CCCCGTCAATTCTATTGAGTTT). The PCR mixtures consisted of a total volume of 25 µL, containing 1 µL of genomic DNA, 2.5 µL of 10× reaction buffer, 2 µL of 2.5 mM deoxy-ribonucleoside triphosphate (dNTP) mix, 1 µL of 10 µM forward primer, 1 µL of 10 µM reverse primer, 0.3 µL of 5 U/µL Taq-T DNA polymerase (TransGen Biotech, Beijing, China), and ddH₂O to 25 µL. The PCR products were detected by electrophoresis in 1% agarose gels.

The PCR products were sequenced by GENEWIZ (GENEWIZ, Beijing, China). The sequences obtained were analyzed using the National Center for Biotechnology Information (NCBI) BLAST program (https://blast.ncbi.nlm.nih.gov/Blast.cgi). Each isolate was compared against known taxa present in the GenBank database.
2.5. Determination of Ligninolytic Enzymatic Activity

Ligninolytic enzymes are key enzymes that act in the degradation of lignin. Therefore, the higher the activity of strains producing laccase, manganese peroxidase, and lignin peroxidase, the stronger the ability of the given strains to degrade lignin.

Three different media were used to assess the ability of the isolates to degrade lignin: (i) sodium lignosulfonate agar medium, 0.2% sodium lignosulfonate, 2.0 g (NH₄)₂SO₄, 1.0 g K₂HPO₄, 1.0 g KH₂PO₄, 0.2 g MgSO₄, 0.1 g CaCl₂, 0.05 g FeSO₄, 0.02 g MnSO₄, 20.0 g agar, and 1 L tap water (pH = 7.0), (ii) aniline blue agar medium, 10.0 g yeast extract, 20.0 g glucose, 20.0 g agar, 0.1 g aniline blue, and 1 L distilled water, and (iii) Remazol brilliant blue agar medium, 10.0 g yeast extract, 20.0 g glucose, 20.0 g agar, 0.1 g Remazol brilliant blue, and 1 L distilled water. All media were autoclaved for 20 min at 121 °C. The chemical structure of lignosulfonate is almost the same as that of natural lignin. Furthermore, the decolorization of aniline blue is related to the production of lignin peroxidase and manganese peroxidase but does not reflect the production of laccase produced by the strain, while laccase can decolor Remazol brilliant blue.

The screening of lignin-degrading bacteria was carried out in two steps. The bacteria were inoculated on sodium lignosulfonate agar plates using the streak plate method and incubated at 28 °C for 5 days. Potential ligninolytic enzymes were initially obtained according to the colony growth conditions. Then, the bacteria were transferred onto Remazol brilliant blue agar plates and aniline blue agar plates and incubated at 28 °C for 4 days. The lignocellulose degradation ability by the bacteria was directly proportional to the size of the decolorized circles. Furthermore, the larger the decolorized circles, the stronger the lignocellulose degrading ability [14].

2.6. Determination of Cellulolytic Enzymatic Activity

The cellulose degradation by bacteria was visualized on plates with cellulose as the sole carbon source and stained by Congo red. Congo red indicates the activity of cellulolytic enzymes—the dye can bind to the polysaccharide substrate of the medium and form transparent zones around the colony.

The purified colonies were separately transferred onto CMC-Na agar plates, which consisted of 15.0 g CMC-Na, 5.0 g NaCl, 1.0 g KH₂PO₄, 0.2 g MgSO₄, 10.0 g peptone, 5.0 g yeast extract, 18.0 g agar, and 1 L distilled water. All media were autoclaved for 20 min at 121 °C. Three plates were made for each strain and incubated at 28 °C for 4 days. After incubation, the media plates were flooded with 1 g/L Congo red for 15–20 min. The dye was discarded, and the plates were flooded with 1 M NaCl for 20 min. Finally, the NaCl solution was discarded [15]. The diameters of the colonies and transparent circles were measured and termed D (diameter of colonies) and H (hydrolysis circle), respectively. The cellulolytic enzymatic activity of each bacterial colony was preliminarily determined according to the ratio of H and D. Furthermore, the larger the H/D value, the stronger the cellulose degrading ability.

2.7. Determination of Biocide Effectiveness by Inhibition Assay

The application of biocides is one of the most effective methods to control microbial deterioration. The selected biocides were Preventol® D7 (Lanxess, Köln, Germany), BIT 20N, P91, and Euxyl® K100 (Schülke, Norderstedt, Germany). All of these chemicals are isothiazolinone derivatives, which are considered broad-range biocides [16].

The purified single colonies were picked up with a sterile cotton swab and coated on LB agar plates. Filter paper was cut to make circles with a 0.7 cm diameter. Five pieces were put on each LB agar plate, four of which were loaded with 10 µL of different biocides (0.5%). The remaining piece was loaded with sterile water as a control. The plates were examined for clear zones after incubation at 28 °C for 3 days. Measuring the diameter of the clear zones indicated the effectiveness of the biocidal agents. Each antimicrobial agent test comprised three replicates. All materials (media, cotton swabs, etc.) were autoclaved for 20 min at 121 °C.
3. Results

3.1. Bacterial Community Analyses

The diversity of the bacterial community was assessed using high-throughput sequencing on an Illumina HiSeq 2500 PE250 platform (Novogene Bioinformatics Technology Co., Ltd. Beijing, China) for the four stored water samples. Two of the samples were collected in June, and another two were collected in December 2017. The ten most-abundant genera are presented in Figure 1. The bacterial communities in the four water samples mostly consisted of Xanthobacter, Rhizobium, Acinetobacter, Variorovax, Flavobacterium, Pseudomonas, Neorhizobium, Fusibacter, Sphingobacterium, Brevundimonas, Geothrix, Novosphingobium, Sphingobium, Cupriavidus, Haliscomenobacter, Microbacterium, Ralstonia, Unidentified_Mitochondria, and others. Others were predominant in all four water samples, and the genera Flavobacterium and Pseudomonas were detected in all four samples. The most abundant identified genus of the two water samples collected in June was Xanthobacter, which accounted for 21.73% of the detected signatures with an average of 10.87% in the two samples. Rhizobium and Acinetobacter took the second and third place, accounting for 14.86% and 11.20% with an average of 7.43% and 5.60%, respectively. Meanwhile, Xanthobacter represented the greatest bacterial proportion of sample No. 9, accounting for 18.46%, and Acinetobacter was the primary genus in sample No. 10, accounting for 3.74%. Other major genera, including Variorovax and Flavobacterium, comprised bacterial signatures in the two water samples of 7.44% and 5.49%, respectively. Flavobacterium was present in the greatest bacterial proportion of the identified genera in the two water samples collected in December, accounting for 16.60% with an average of 8.30%. Pseudomonas and Novosphingobium followed in proportion, accounting for 7.21% and 6.87% with an average of 3.61% and 3.44%, respectively. The dominant genera of the two water samples were Flavobacterium (No. NHWS03) and Novosphingobium (No. NHWS04), accounting for 12.81% and 5.24%, respectively. In addition, Brevundimonas, Geothrix, Microbacterium, and Haliscomenobacter were other major members of the bacterial communities in the two samples, and they represented 6.00%, 5.35%, 5.31%, and 4.70%, respectively.

**Figure 1.** Relative abundance of the ten most-abundant microbial genera among the four samples. The relative abundance for each sample is shown as a percentage. Bacterial genera are colored according to the legend on the right. (A) Shows water samples No. 9 and No. 10 collected in June 2017, and (B) shows water samples No. NHWS03 and No. NHWS04 collected in December 2017.

The distribution of the bacterial communities in the two collections was different; however, the two samples collected at the same time had similar bacterial communities. Xanthobacter, Rhizobium, and Acinetobacter were the three most-abundant genera in the collection sampled in June, but these were not present in the top ten genera of the samples collected in December. In the same way, the genus Novosphingobium, the third largest member of the bacteria represented in the samples collected in December, was present as one of the ten most-abundant genera in June. The composition of other major bacterial communities was also quite different.
3.2. Isolation of Dominant Bacteria

Two water samples were smeared on LB medium, M2 medium, and MacConkey agar (MAC) medium and incubated for 4–5 days. Then, different colonies were selected, purified, and isolated by cross lining on the LB medium. After DNA molecular identification, the isolated and pure strains were named *Microbacterium* sp. NK-NH4 (MH997669), *Bacillus tequilensis* NK-NH5 (MH997670), *Bacillus subtilis* NK-NH6 (MH997671), *Pseudomonas* sp. NK-NH7 (MH997672), *Bacillus* sp. NK-NH8 (MH997673), *Ochrobactrum* sp. NK-NH9 (MH997674), *Bacillus megaterium* NK-NH10 (MH997675), *Bacillus velezensis* NK-NH11 (MH997676), *Microbacterium* sp. NK-NH12 (MH997677), and *Stenotrophomonas* sp. NK-NH14 (MH997679). The results are presented in Table 1.

<table>
<thead>
<tr>
<th>Bacteria Accession Number</th>
<th>Closet Relative Strain Accession Number</th>
<th>Similarity (%)</th>
<th>Source</th>
</tr>
</thead>
<tbody>
<tr>
<td>NK-NH4 MH997669</td>
<td><em>Microbacterium</em> sp. DQ658916.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH5 MH997670</td>
<td><em>Bacillus tequilensis</em> MG519475.1</td>
<td>93%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH6 MH997671</td>
<td><em>Bacillus subtilis</em> K510927.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH7 MH997672</td>
<td><em>Pseudomonas</em> sp. MF436694.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH8 MH997673</td>
<td><em>Bacillus</em> sp. MG827113.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH9 MH997674</td>
<td><em>Ochrobactrum</em> sp. MF442305.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH10 MH997675</td>
<td><em>Bacillus megaterium</em> MG818943.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH11 MH997676</td>
<td><em>Bacillus velezensis</em> MH014956.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
<tr>
<td>NK-NH14 MH997679</td>
<td><em>Stenotrophomonas</em> sp. MH144251.1</td>
<td>99%</td>
<td>water samples</td>
</tr>
</tbody>
</table>

3.3. Degradation of Cellulose and Lignin by Isolated Bacterial Strains

All isolated strains were cultured on a medium containing carboxymethylcellulose (CMC) to determine whether they could degrade cellulose. These strains were cultured on CMC plates at 28 °C for 4 days (Shown in Figure S2). As shown in Figure 2, after flooding the plates with Congo red for 15–20 min and eluting with NaCl solution for 20 min, *Microbacterium* sp. NK-NH4, *Ochrobactrum* sp. NK-NH9, and *Bacillus megaterium* NK-NH10 barely grew and almost did not show degradation circles. The rest of the cultured bacteria showed different sizes of degradation circles. *Bacillus tequilensis* NK-NH5 and *Bacillus velezensis* NK-NH11 had the highest capacity for cellulose degradation (H/D > 3.0). *Stenotrophomonas* sp. NK-NH14 also showed a high capacity for cellulose degradation (H/D > 2.0). *Bacillus subtilis* NK-NH6, *Pseudomonas* sp. NK-NH7, and *Bacillus* sp. NK-NH8 had a relatively lower capacity for cellulose degradation (H/D < 2.0), even though these genera grew more vigorously on the plates.
Bacillus megaterium (H/D  < 2.0). Among them, Microbacterium sp. NK-NH8, Bacillus subtilis NK-NH5, and Stenotrophomonas sp. NK-NH9 barely grew. Bacillus tequilensis NK-NH10 showed degradation zones; however, these zones were not clear, and the ratio of H and D was lower than 1.5.

As shown in Figure 3, on the medium with Remazol brilliant blue, Pseudomonas sp. NK-NH7, Bacillus sp. NK-NH8, Ochrobactrum sp. NK-NH9, and Stenotrophomonas sp. NK-NH14 did not show degradation zones. The rest of the cultured bacteria showed different sizes of decolorized circles, indicating that they had the ability to produce laccase; however, the enzyme production was weak (H/D < 2.0). Bacillus megaterium NK-NH10 had the most distinct degradation circle and presented a relatively high ability to produce laccase together with Microbacterium sp. NK-NH4. Bacillus tequilensis NK-NH5, Bacillus subtilis NK-NH6, and Bacillus velezensis NK-NH11 showed degradation zones; however, these zones were not clear, and the ratio of H and D was lower than 1.5.
were different. Among the four agents, the efficacy of Euxyl® K100 was the smallest, while Preventol® D7 and BIT 20N, P91, and Euxyl® K100, which can inhibit bacterial growth. All the isolated strains showed inhibition rings at a 0.5% concentration of biocide, indicating that the strains were all sensitive to isothiazolinones. However, the effects exerted by the different agents on bacterial strains showed inhibition rings at a 0.5% concentration of biocide, indicating that the strains were all sensitive to isothiazolinones. However, the effects exerted by the different agents on bacterial strains varied with the various strains. The results are presented in Figure 4 and Table S1.

3.4. Biocide Susceptibility

To observe the susceptibility of the strains to biocides, we measured the strain growth by adding four kinds of biocides to LB plates. The drugs containing isothiazolinones included Preventol® D7, BIT 20N, P91, and Euxyl® K100, which can inhibit bacterial growth. All the isolated strains showed inhibition rings at a 0.5% concentration of biocide, indicating that the strains were all sensitive to isothiazolinones. However, the effects exerted by the different agents on bacterial growth suppression were different. Among the four agents, the efficacy of Euxyl® K100 was the smallest, while Preventol® D7 and BIT 20N were the most effective. However, the efficacy of the individual agents varied with the various strains. The results are presented in Figure 4 and Table S1.
4. Discussion

Microorganisms play an important role in the degradation of woody artifacts [17]. Therefore, the inhibition of microbial corrosion is critical in maintaining important cultural heritage projects. Given that bacteria are the main degraders of archaeological wood under near-anaerobic conditions [18], we focused on the bacterial communities present in the water used to store the lacquer plates from an excavated 800-year-old shipwreck.

High-throughput sequencing results showed that the most abundant bacteria in the water samples were *Flavobacterium*, *Pseudomonas*, *Xanthobacter*, *Rhizobium*, *Acinetobacter*, and *Novosphingobium*. Among them, *Flavobacterium* and *Pseudomonas* were detected in all four water samples. After traditional microbial culture, isolation, purification, and 16s rRNA gene identification, *Pseudomonas, Microbacterium, Bacillus, Ochrobactrum*, and *Achromobacter* were obtained.

*Pseudomonas* is an obligate aerobic gram-negative genus and is common in a variety of terrestrial environments. Recently, *Pseudomonas* has been found in Antarctic soil and even in deep sea sediments [19]. *Microbacterium* is a genus of gram-positive aerobic bacteria. Studies have shown that *Microbacterium* may have the ability to degrade lignin [20]. *Bacillus* is generally an aerobic heterotrophic bacterium, and the presence of *Bacillus* indicates that the environment is rich in organic matter. Gontang et al. found that *Bacillus licheniformis* is a type of cultivable bacteria that can be found in the 0- to 500-meter intertidal zone sediments of the Western Pacific Islands (7°30’N, 134°30’E) [21]. Sun et al. determined the diversity and abundance of *Bacillus* in South China Sea sediments, where *Pseudomonas, Microbacterium, Achromobacter*, and other genera are present [12]. *Ochrobactrum* is a genus of gram-negative bacteria with a strict respiratory metabolism that is obligate aerobic. Bao et al. isolated an *Ochrobactrum* sp. LJY313 from sea mud samples from the Arctic Ocean [22]. *Flavobacterium* is a genus of gram-negative, nonmotile and motile bacteria. *Flavobacteria* are found in soil and fresh water in a variety of environments [23]. *Xanthobacter* is an aerobic, spherical- to rod-shaped gram-negative *Bacillus* of unidentified genus. This genus consists of inorganic or organic nitrogen-fixing bacteria that live in moist soils and waters. *Rhizobium* is a genus of gram-negative, motile, nonsporulating rods. *Rhizobium* has nitrogen fixation abilities but has not yet been found in the ocean. The species of the genus *Acinetobacter* are strictly aerobic, nonfermentative, gram-negative bacilli. *Acinetobacter* species are widely distributed in nature and commonly occur in soil and water. They are able to survive on moist and dry surfaces and have the capacity to survive exposure to various common disinfectants. Furthermore, *Acinetobacter* species can grow at a broad range of temperatures, allowing them to survive in a broad array of environments [24]. *Novosphingobium* is a genus of gram-negative bacteria. *Novosphingobium* is ubiquitous in the environment, and there are traces of it in rivers, rhizospheres, surface sediments, deep underground sediments, oceans, and even polar soils. *Achromobacter* is a genus of gram-negative bacteria; *Achromobacter* species are strictly aerobic and are found in water (fresh and marine) and soils [25].

Almost all bacteria of known physiological groups can be found in marine environments. Studies have found that *Xanthobacter, Novosphingobium, Pseudomonas, Microbacterium, Flavobacterium*, and *Bacillus* appear in eutrophic lakes [26]. Bacteria, such as *Pseudomonas, Flavobacterium, and Bacillus*, have been reported to have extensive algae-dissolving abilities [27,28].

In fact, biological degradation is the main threat to archaeological wood [29]. It is also impossible to ignore the influence of bacteria living in the water on waterlogged archaeological wood [30]. There are indeed a number of wooden artifacts suffering from the biodegradation caused by bacteria.

According to M. D. Herraiz and V. Jurado, several types of dominant bacteria, such as *Actinobacteria, Alphaproteobacteria, Gammaproteobacteria, and Acidobacteria*, were isolated from the samples collected from the Tomba della Scimmia (Tomb of the Monkey) 480–470 BC near Chiusi, Italy [31]. According to C. G. Björdl, T. Nilson, and G. Daniel, the microbial decay of waterlogged archaeological wood found in Sweden was mainly caused by bacterial corrosion [32]. A similar case is also exemplified by the Xiao bajiaio shipwreck (1821–1850) in Zhejiang province [9]. Moreover, according to Y. H. Chen, bamboo slips, belonging to the Three Kingdoms period, that were excavated...
from Changsha, Hunan have suffered from degradation by seven genera of bacteria: *Bacillus*, *Acinetobacter*, *Staphylococcus*, *Pandoraea*, *Novosphingobium*, *Cupriavidus*, and *Comamonas* [8].

Microbial interactions also have important effects on the degradation of cellulose and lignin in wood. Studies on the biodegradation of cellulose have shown that the degradation of cellulose by a single microorganism is difficult or only weak and must rely on the combined action of two or more microorganisms [33]. Due to the synergistic effect and interdependence of different species in the microbial community, it is speculated that the interaction of microorganisms in the liquid matrix may greatly improve enzyme activity and promote the degradation of cellulose. The biodegradation of lignin under natural conditions is also the result of the combined action of many kinds of microorganisms, among which the role of fungi is the most important. For example, the high-throughput sequencing of fungi involved in our research was conducted, though the results were not published; *Penicillium* sp. and *Fusarium* sp. were the most predominant fungi. A rising number of studies on the bacterial degradation of lignin have proved that bacteria can degrade lignin in a liquid matrix and, at the same time, have a good ability to adapt and regulate the environment [34]. Bacteria degrade lignin in the primary metabolic stage and synthesize lignin-degrading enzymes in the logarithmic and stable growth stages [35]. In the degradation process, bacteria mainly denaturate lignin to a certain extent and make it a water-soluble polymer, but they seldom mineralize lignin into CO$_2$. In addition, bacteria mainly play an indirect role in the degradation of lignin, and there are synergistic effects between bacteria and fungi in the degradation of lignin [36].

For waterlogged wooden artifacts, we suggest that some effective measures can be taken to conserve them. Most wooden remains from archaeological sites in waterlogged burial environments are found to be intact without shrinkage, despite centuries of storage. However, waterlogged wooden remains are subject to shrinkage damage, and some even lose up to 80% of their original volume when exposed to air [37]. When such objects are allowed to dry, they may collapse [38]. This is the reason why lacquer plates excavated from the Nanhai No. 1 shipwreck need to be conserved in water prior to reinforcement. Hence, it has been feasible to store the lacquer plates in sterilized distilled water or double distilled water (ddH$_2$O), providing an environment with few microorganisms. While in storage, the water should be replaced at regular intervals. Furthermore, cold, dark and near-anaerobic conditions result in slower degradation [38]. Since the lacquer plates from the Nanhai No. 1 shipwreck involved in this research still need to be stored in water, the measure of maintaining an anaerobic environment needs to be considered. Meanwhile, other waterlogged wooden artifacts may take this method into account. In addition, the temperature used for wooden storage should be taken into consideration—waterlogged wood should be preserved at a low temperature of approximately 4°C to inhibit bacterial growth.

In addition to controlling the preservation environment, undertaking efficient monitoring and using protective measures are imperative, as the results of the high-throughput sequencing showed that the change in bacterial communities at different sampling times was remarkable. First, the water samples used to store lacquer plates should be constantly monitored to assess whether the microbial communities change. Second, the application of efficient and low-toxicity biocides may be considered to restrain bacterial growth and mitigate biodeterioration [39]. Although the biocide susceptibility test illustrated the great capacity of biocides to chemically repress the growth of prominent bacteria, biocides can exert pressure on microbial communities, resulting in the development of biocide-resistant mechanisms or the replacement of dominant bacteria with new microbial communities, which may bring about more serious damage to cultural artifacts [2]. For example, in the French Lascaux Cave, a series of biocide treatments were applied that led to new microbial outbreaks [40]. Therefore, the application of biocides in moderation may be the optional approach. Before using biocides, however, their potential harm to humans, the environment, and to cultural relics should be taken into account, and further tests may be required to evaluate these factors.

**Supplementary Materials:** The following materials are available online at http://www.mdpi.com/2076-3417/9/4/653/s1: Figure S1—the form of lacquer plates soaking in the water; Figure S2—colonies of nine strains
grown on azure plates at 28 °C for 4 days; and Table S1—describes the effectivity of different biocides against different strains.

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