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Estimation of Forest NPP and Carbon Sequestration in the Three Gorges Reservoir Area, Using the Biome-BGC Model

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Abstract: The Three Gorges Reservoir area is one of the most ecologically sensitive areas in China, and the forest landscape pattern in this region shows dramatic change due to the influence of the Three Gorges reservoir project. In this study, the locally parameterized Biome-BGC model, generated with long-term meteorological monitoring data, was used to simulate net primary productivity (NPP) and carbon density of the vegetation layer, the litter layer, and the soil layer for various forest types from 1992 to 2012 in this area. The total and unitary forest NPP presented obvious annual fluctuation under the combined influences of land use change and extreme weather events. Apart from the year 2006, from 1992 to 2012, the NPP values of each forest type showed an increasing trend, although the growth rates decreased. In 2006, due to abnormally high air temperatures and less precipitation, total and unit area forest NPP values decreased by 46.3% and 53.9%, respectively, compared to 2002. From 1992 to 2012, the carbon stocks of the forest vegetation layer, the litter layer, the soil layer, and the entire area gradually increased with decreasing growth rates. Additionally, forest carbon stocks were high in the east and the south and low in the west and the north. Generally, the forest productivity is greatly affected by the physiological and ecological characteristics of the plants themselves as well as the environmental factors, whereas total forest productivity is largely influenced by human activities. The increase in forest area and the optimization of the forest landscape pattern could improve the forest productivity and carbon sequestration.

Keywords: forest NPP; forest carbon sequestration; Biome-BGC model; Three Gorges Reservoir area

1. Introduction

Climate warming is currently receiving considerable attention and has become a global research focus. Forests are the main vegetation cover in terrestrial ecosystems, and forest productivity and biomass play an important role in global climate change and circulation of substances [1]. Via photosynthesis, forests fix atmospheric CO₂ and accumulate carbon in their biomass, litter, and soil. In addition, due to biological respiration, litter decomposition, soil carbon degradation, and natural or artificial disturbances, forests emit carbon into the atmosphere in the form of CO₂. Although the forest area only accounts for about one third of the total terrestrial

ecosystem area, carbon storage in aboveground forest vegetation is equivalent to 80% of the aboveground total organic carbon of terrestrial ecosystems; belowground carbon storage accounts for 40% of that of the entire terrestrial ecosystem [2].

The Three Gorges Reservoir area (TGRA) of the Yangtze River is a new concept derived from the construction of the Three Gorges Dam Project since 1992. It refers to the region submerged due to the water level rise caused by the Three Gorges Dam (water level 175 m; height of the dam wall 185 m). The Three Gorges Reservoir Area (TGRA) of the Yangtze River is the most ecologically sensitive and fragile area in China with complex terrain and frequent natural disasters, which is located in the transition zone between a mountainous region and watershed region. As a special human activity disturbance, the Land Use/Cover Change (LUCC) was seriously affected by the construction of the reservoir, including the establishment of new towns and cities, migration of millions of people and the extended railway and highway nets [3]. In recent decades, the built-up land increased the most, largely and rapidly, followed by water bodies. They were transformed from croplands, forests and grasslands.

Against the background of climate change and land use change, the present situation of the forest ecosystem in the Three Gorges Reservoir area, as well as its response to global climate change, has become the focus of a large number of studies both in China and around the world [4–6]. This study focused on the temporal and spatial variations in forest productivity and carbon storage in the TGRA as well as the corresponding influencing factors. The research results are significant in revealing the extent and scope of human disturbance effects on forest productivity and forest carbon storage and to provide evidence for predicting the response of forest productivity to future climate change scenarios. Therefore, a locally parameterized Biome-BGC model was used to simulate net primary productivity (NPP) and carbon density of the vegetation layer, the litter layer, and the soil layer for various forest types from 1992 to 2012 in the TGRA in order to reveal the changes.

2. Materials and Methods

2.1. Study Area

The Three Gorges Reservoir Area is located at $106^{\circ}16'-111^{\circ}28'E$, and $28^{\circ}56'-31^{\circ}44'N$ and covers a total of 26 district (counties), 22 of which are located in Chongqing Municipality directly under the Central Government while 4 are situated in Hubei Province. The reservoir catchment covers an area of about 5.8×10^4 km², with a total population of about 22 million people. The climate is humid subtropical monsoon, with an annual average temperature of 16–18 °C and an annual precipitation of 1000–1300 mm, with early springs, hot summers, wet autumns, and warm winters [7].

The TGRA is situated on the eastern margin of the second step of the three geomorphic steps in China. It is at the junction of three tectonic units: the Daba Mountain fold belt, the parallel ridge valley in eastern Sichuan and the Sichuan-Hubei-Hunan-Guizhou uplift fold belt. Mountains account for about 74%, hills account for about 22%, plains and dams account for only 4%. The terrain of this area is high in the east and low in the west with an altitude range from 50 meters to 2900 meters. From the east to the west, the area can be divided into three parts: the eastern mountain part, the central parallel ridge valley part and the western hills (Figure 1).

The TGRA has rich plant resources. It contains 6088 species of vascular plants, belonging to 1428 genera and 208 families, and the species in this area account for about 20% of the total plant species in China [8]. The main species are Pinaceae、 Cupressaceae、 Juglandaceae、 Fagaceae、 Hamamelidaceae、 Roseceae、 Rutaceae、 Anarcardiaceae、 Theaceae、 Cornaceae、 Ericaceae、 Betulaceae、 Leguminosae、 Gramineae and Pteridophyta [9,10]. Based on the characteristics of the forest types and the main species composition, the forest vegetation in the TGRA can be divided into four major types, namely, coniferous forest, deciduous broadleaf forest, evergreen broadleaf forest, and shrub forest [11,12]. Due to excessive exploitation and interference for decades, the natural forest area is extremely rare in this region, and most of the natural forests are secondary forests. The forest type structure is relatively simple, and *Pinus massoniana* and *Cupressus funebris* are the most

widely distributed species with the highest abundance [13]. The forests are mainly distributed in the eastern mountain part, the western Karst Mountains, and the low hilly areas in the south (Figure 2).

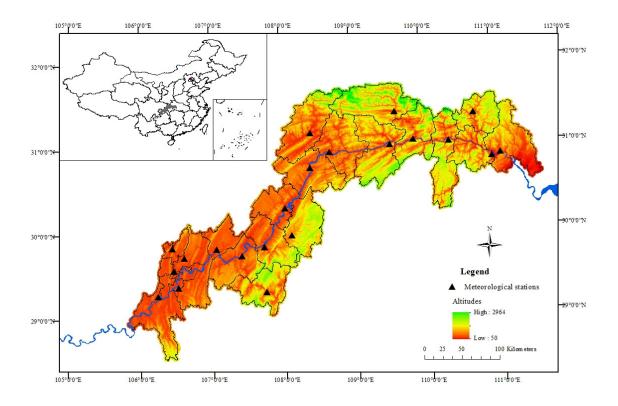


Figure 1. Location and administrative districts of the TGRA.

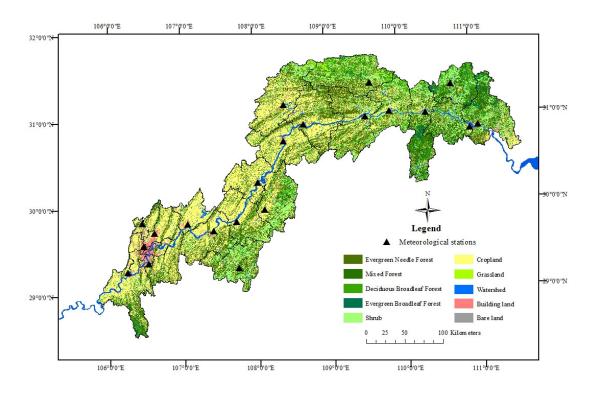


Figure 2. The distribution map of forest types in the TGRA in 2012.

[14,34].

The Biome-BGC model is one of the typical biogeochemical models. The Biome-BGC version 4.2 (Numerical Terradynamic Simulation Group, Missoula, MT, USA) was used in this study, which can predict the flow and storage of carbon, nitrogen, water, and radiation in terrestrial vegetation and soil, with a time-step of one day, based on daily meteorological data. In addition, the Biome-BGC model considers how global climate change and circulation of substances are affected by environmental conditions such as temperature, precipitation, solar radiation, soil texture, and atmospheric CO₂ concentration [14–17]. The Biome-BGC model has been widely used. For example, Ueyama M et al. applied this model to simulate the carbon flux of an Alaska black pine forest and verified the model by comparing the simulated values with the measured values by the eddy covariance method. Their results showed that gross primary productivity (GPP), net ecosystem carbon exchange capacity (NEE), and ecosystem respiration rates (RE) obtained by the two methods were highly correlated [18]. Similarly, they also found in another study that the GPP, NEE, and RE values simulated by the Biome-BGC model were highly consistent with the observed values based on AsiaFlux [19]. In China, this model has also been used to simulate the characteristics of NPP under current climatic conditions and to predict the NPP characteristics under future climate change scenarios, not only for a particular tree species, but also for forest ecosystems in different regions, such as *Picea schrenkiana* in Tianshan Mountains of Xinjiang [20], a *Pinus elliottii* plantation in a hilly Krasnozem area [21], a Quercus variabilis forest in the Beijing mountainous area [22], a Quercus variabilis forest [23], a Larix principis-rupprechtii forest in Beijing mountain area [24], a Robinia pseudoacacia forest [25], as well as the forest ecosystem in Fujian Province located in a subtropical zone and which has the highest forest coverage in China [26], and a forest ecosystem on the northern slope of Tianshan Mountains in the arid region of China[27]. Compared with and verified by the measured values estimated according to the classical ecological survey method, the simulated values had a high correlation. Therefore, the Biome-BGC model can not only reproduce dynamic characteristics of short-term carbon fluxes but can also simulate long-term distribution patterns and dynamics of carbon storage as well as the responses of these processes to environmental changes, including climate change [28-33].

2.3. Parameterization of the Biome-BGC Model

Three files are required to run the Biome-BGC model: initialization file, meteorological data file, ecophysiological constants file.

Firstly, the Biome-BGC model was spun-up based on a set of long-term meteorological data, site data, and ecophysiological data. As a hypothesis, the CO₂ concentration values and nitrogen deposition values prior to the industrial revolution were used as a steady state ecosystem. After repeated and continuous simulation, the model reached an equilibrium state. In the state of equilibrium, the meteorological data, site data, and ecophysiological data during the study period were launched into the Biome-BGC in order to run the model and then obtain the final results

Initialization files are used to control the running process of the model. Firstly, the reading paths of the prepared meteorological and ecophysiological parameters files are specified. Then, the environmental information of each county needs to be described (Table 1), including latitude, altitude, effective depth of soil, soil texture, albedo, etc. Finally, the result variables of model simulation output are defined. The soil characteristic parameters in this study were based on the "National 1:1 million Digital Soil Map" developed by Nanjing Soil Research Institute of Chinese Academy of Sciences and Soil Environment Department of the Ministry of Agriculture of China.

No	County	Latitude	Longitude	Altitude (m)	Soil type	Annual Precipitation (cm)
1	Fengjie	31°01′N	109°32′E	607.3	Yellow rendzina	112.79
2	Badong	31°02′N	110° 22′ E	334.0	Dark yellow-brown soil	109.04
3	Zigui	30°50′N	110°58′E	295.5	Yellow soil	115.86
4	Xingshan	31°21′N	110°44′E	336.8	Yellow soil	101.59
5	Wanzhou	30°46′N	108°24′E	186.7	Neutral purple soil	124.52
6	Yiling	30°42′N	111°18′E	133.1	Yellow soil	122.41
7	Chongqing	29°35′N	106°28′E	259.1	Neutral purple soil	113.58
8	Jiangjin	29°17′N	106°15′E	261.4	Yellow soil	102.41
9	Changshou	29°50′N	107°04′E	377.6	Neutral purple soil	115.59
10	Fengdu	29°51′N	107°44′E	290.5	Neutral purple soil	104.20
11	Kaixian	31°11′N	108°25′E	216.5	Calcareous purple soil	126.01
12	Yunyang	30°57′N	108°41′E	297.2	Acid purple soil	110.78
13	Wuxi	31°24′N	109°37′E	337.8	Yellow soil	109.05
14	Wushan	31°04′N	109°52′E	275.7	Yellow rendzina	102.96
15	Zhongxian	30°18′N	108°02′E	325.6	Calcareous purple soil	119.50
16	Shizhu	29°59′N	108°07′E	632.3	Yellow soil	106.63
17	Yubei	29°44′N	106°37′E	464.7	Neutral purple soil	115.56
18	Banan	29°23′N	106°32′E	243.6	Yellow soil	107.02
19	Fuling	29°45′N	107°25′E	273.5	Neutral purple soil	110.69
20	Wulong	29°19′N	107°45′E	277.9	Yellow soil	102.49

Table 1. The environmental information of the counties in the TGRA.

The meteorological data file is comprised of daily maximum temperature, daily minimum temperature, daily average temperature, daily precipitation, daily water vapor pressure difference, daily short wave radiation, and day length. Measured values from the national ground stations of China Meteorological Administration within the TGRA were used; each district (county) had a located meteorological station (Figures 1 and 2 shown the distribution of the stations), with a total of 34 years (from 1981 to 2014) of observations.

The ecophysiological parameter files of the Biome-BGC model are based on seven types: deciduous broadleaf forest (DBF), evergreen broadleaf forest (EBF), evergreen needle leaf forest (ENF), deciduous needle leaf forest (DNF), shrub forest (SHRUB), C3 grassland, and C4 grassland. The Numerical Terradynamic Simulation Group (NTSG) of the Montana University provided a set of default ecophysiological parameters, containing 42 parameters for each vegetation type. These default values are assigned based on the average value of parameters obtained from the literature research or the value determined by clustering analysis after literature retrieval and statistical analysis of each parameter for each community [35]. According to the forest vegetation types and the main species in the TGAR, the ecophysiological constants of Pinus massoniana and oaks were used for the ENF and DBF, while the default parameters were used for the EBF and SHRUB. For the mixed coniferous broadleaved forest (MIX), the Biome-BGC model does not provide the corresponding ecophysiological parameters, making a direct assessment of these parameters impossible. However, previous studies have shown that we can effectively simulate the productivity of MIX forests by simulating the productivity of both vegetation types respectively and using a subsequent proportion conversion like 50% for each forest type [36,37]. Some parameters, such as carbon allocation from roots to leaves, carbon allocation from stems to leaves, carbon allocation from thick roots to stems, carbon to nitrogen ratio of leaves, carbon to nitrogen ratio of leaf litter, carbon to nitrogen ratio of roots, and carbon to nitrogen ratio of woody parts, can be locally parameterized by actual measurement or literature findings [22,23,26,38–40]; for other parameters that are difficult to be obtained, default values can be adopted, which are provided by the model (Table 2).

Table 2. Value of ecophysiological parameters of forests used in Biome-BGC.

No	Parameters	Unit	ENF	DBF	EBF	SHRUB
P1	Annual leaf and fine root turnover fraction	year-1	0.25	1.0	0.5	0.25
P2	Annual live wood turnover fraction	year ⁻¹	0.7	0.7	0.7	0.7
Р3	Annual whole-plant mortality fraction	year ⁻¹	0.005	0.005	0.005	0.02
P4	Annual fire mortality fraction	year ⁻¹	0.005	0.0025	0.002	0.01
P5	New fine root C : new leaf C	ratio	0.8	0.563	1.0	0.815
P6	New stem C : new leaf C	ratio	2.2	0.67	1.0	0.5
P7	New live wood C : new total wood C	ratio	0.071	0.16	0.22	1
$\mathbf{P8}$	New croot C : new stem C	ratio	0.29	0.77	0.3	0.655
Р9	Current growth proportion	prop.	0.5	0.5	0.5	0.5
P10	C:N of leaves	KgC/KgN	35.0	35.9	42.0	33.425
P11	C:N of leaf litter, after retranslocation	KgC/KgN	93.0	65.72	49.0	75
P12	C:N of fine roots	KgC/KgN	58.0	43.23	42.0	52.515
P13	C:N of live wood	KgC/KgN	58.0	47.23	50.0	58
P14	C:N of dead wood	KgC/KgN	729.0	109.9	300.0	730
P15	Leaf litter labile proportion	DIM	0.31	0.39	0.32	0.48
P16	Leaf litter cellulose proportion	DIM	0.45	0.44	0.44	0.37
P17	Leaf litter lignin proportion	DIM	0.24	0.17	0.24	0.15
P18	Fine root labile proportion	DIM	0.34	0.29	0.30	0.41
P19	Fine root cellulose proportion	DIM	0.44	0.18	0.45	0.37
P20	Fine root lignin proportion	DIM	0.22	0.53	0.25	0.22
P21	Dead wood cellulose proportion	DIM	0.71	0.66	0.76	0.71
P22	Dead wood lignin proportion	DIM	0.29	0.34	0.24	0.29
P23	Canopy water interception coefficient	1/LAI/day	0.045	0.021	0.041	0.045
P24	Canopy light extinction coefficient	DIM	0.5	0.7	0.7	0.55
P25	All-sided to projected leaf area ratio	DIM	2.6	2.0	2.0	2.3
P26	Specific leaf area	m²/KgC	8.1	20.0	12.0	17.5
P27	Ratio of shaded SLA:sunlit SLA	DIM	2.0	2.0	2.0	2
P28	Fraction of leaf N in Rubisco	DIM	0.08	0.085	0.06	0.04
P29	Maximum stomatal conductance	m/s	0.006	0.005	0.005	0.006
P30	Cuticular conductance	m/s	0.00006	0.00001	0.00001	0.00006
P31	Boundary layer conductance	m/s	0.09	0.01	0.01	0.02
P32	Leaf water potential: start of conductance reduction	MPa	-0.65	-0.6	-0.6	-0.81
P33	Leaf water potential: complete conductance reduction	MPa	-2.5	-2.3	-3.9	-4.2
P34	Vapor pressure deficit: start of conductance reduction	Pa	610	930.0	1800.0	970
P35	Vapor pressure deficit: complete conductance reduction	Ра	3100	4100	4100.0	4100

2.4. Validation of The Biome-BGC Model

In this study, the simulation results of the Biome-BGC model were validated through comparative analysis of observed values and simulated values, combined with relevant literature findings. The observed values of NPP were derived from the analytic-tree fitted biomass model, obtained from a sample-plot survey based on the sub-compartment data of forest resource inventory and planning projects in 2003 and 2013 [41,42]. The carbon content in various plant organs, litters, and A and B layers of soil are determined by experiment, and the observed values of carbon density were estimated according to the carbon storage obtained from the biomass and the area data [43]. Fu et al. used the CBM-CFS3 model as the basic platform and simulated annual NPP values (2.20–5.08 Mg C hm⁻² year⁻¹) and carbon densities of various forest types in the Three Gorges Reservoir area by plotting age-harvest growth curves for the major forest types [44]. Wang et al., based on forest resources inventory data from three periods (6, 7, and 8), used the forest stock volume (increment) extension method to estimate average NPP values of forests in the Three Gorges Reservoir area; NPP

ranged between 2.85 and 6.19 Mg C hm⁻² year⁻¹ [45]. The comparison results are shown in Tables 3 and 4.

Vegetation Type	Observed NPP value (Mg C hm ⁻² year ⁻¹)	Simulated NPP value (Mg C hm ⁻² year ⁻¹)	NPP values from the literature (Mg C hm ⁻² year ⁻¹)
ENF	3.60	5.53	American NIPD of mariana famout
Mixed	3.22	4.94	Average NPP of various forest
DBF	3.28	4.35	types
EBF	4.55	4.41	2.20–5.08 [44]
Shrub	2.81	2.86	2.85–6.19 [45]

Table 3. Comparison between measured and simulated NPP values in the TGRA.

Forest	Vegetation carbon density (Mg hm ⁻²)		Litter carbon density (Mg hm ⁻²)		Soil carbon density (Mg hm ⁻²)			Total carbon density (Mg hm ⁻²)				
Forest type	Biome-BGC	CBM-CFS3	Observed values	Biome-BGC	CBM-CFS3	Observed values	Biome-BGC	CBM-CFS3	Observed values	Biome-BGC	CBM-CFS3	Observed values
ENF	35.04	31.94	24.04	4.27	12.75	2.87	65.45	47.81	98.40	104.76	92.50	125.31
Mixed	31.35	36.01	28.85	3.86	25.72	2.28	65.65	76.49	71.10	100.87	138.21	102.23
DBF	27.67	33.03	26.20	3.46	18.52	2.99	65.85	61.27	96.70	96.98	112.82	125.89
EBF	26.80	41.84	39.71	2.87	17.40	2.94	63.97	58.82	67.60	93.65	118.06	110.25
Shrub	11.51		8.67	1.44		1.30	67.30		82.90	80.25		92.87

Table 4. Comparison of estimation results of carbon density for the main forest types in the TGRA.

By comparison, the NPP values of the ENF and the Mixed coniferous and broadleaf forests, simulated by the Biome-BGC model, were slightly larger than the sample-plot measured values. The simulated carbon density values of the ENF and the Mixed coniferous and broad-leaf forest were close to the measured values, whereas the simulated carbon density values of the DBF, the EBF, and the Shrub forest were lower than the measured values. In spite of this, the values simulated via the Biome-BGC model were consistent with the values simulated by the CBM-CFS3 model and agreed with the estimation results based on the forest resource inventory. Considering that our research focused on the regional scale, the simulation results based on the Biome-BGC model can accurately reflect the real situation, verifying that this model is suitable for the Three Gorges Reservoir area.

2.5. Data Processing and Analysis

According to the different construction stages of the Three Gorges project, we selected five research time-points, namely, 1992, 1996, 2002, 2006, and 2012, to represent five stages of the project process, including the evaluation stage, main channel cutoff, formal impoundment, construction completion, and impoundment to 175 m for full operation, respectively. On this basis, the simulation results from the BIOME-BGC model together with the areas of different forest types derived from remote sensing images at research time-points with spatial geography information were used to analyze the dynamic changes in forest productivity and carbon storage in the reservoir area.

The total forest net primary productivity (NPP) for each respective year was obtained by calculating the NPP values from the Biome-BGC model and the corresponding areas of various forest types for each district (county).

$$NPP_{total} = \sum_{i=1}^{5} \sum_{j=1}^{20} NPP_{i,j} \times S_{i,j}$$
(1)

$$NPP_{mean} = \frac{NPP_{total}}{\sum_{i=1}^{5} \sum_{j=1}^{20} S_{i,j}}$$
(2)

where NPP_{total} is the total forest NPP in a certain year (Tg C); $NPP_{i,j}$ and $S_{i,j}$ represent the NPP values (t hm⁻² year⁻¹) and area (hm⁻²) of the *i*-th (*i* = 1,2,3,...,5) forest type in the *j*-th (*j* = 1,2,3,...,20) district (county) in this year; NPP_{mean} is the area-weighted average NPP value (t hm⁻² year⁻¹) of the forest in the reservoir area in the respective year.

The carbon stocks of forest vegetation, litter, and soil can be calculated by the carbon density values of the vegetation layer, the litter layer, and the soil layer obtained by the Biome-BGC and the corresponding area of various forest types for the district (county) in the respective year. Subsequently, the sum of the above three carbon stocks represents the total forest carbon stock.

$$C_{veg} = \sum_{i=1}^{5} \sum_{j=1}^{20} D_{vegi,j} \times S_{i,j}$$
(3)

$$C_{veg} = \sum_{i=1}^{5} \sum_{j=1}^{20} D_{vegi,j} \times S_{i,j}$$
(4)

$$C_{soil} = \sum_{i=1}^{5} \sum_{j=1}^{20} D_{soili,j} \times S_{i,j}$$
(5)

$$C_{total} = C_{veg} + C_{litr} + C_{soil} \tag{6}$$

where C_{veg} , C_{litr} , and C_{soil} are the carbon stocks of the vegetation layer, the litter layer, and the soil layer in a given year, respectively (Tg C); $D_{vegi,j}$, $D_{litri,j}$, and $D_{soili,j}$ are the carbon densities (t hm⁻²) of the forest vegetation layer, the litter layer, and the soil layer for the *i*-th (*i* = 1,2,3,...,5) forest type in the *j*-th (*j* = 1,2,3,...,20) district (county) in a given year, respectively; $S_{i,j}$ (hm⁻²) is the current area of the

i-th (i = 1,2,3,...,5) forest type in the *j-th* (j = 1,2,3,...,20) district (county); *C*_{total} is the total forest carbon stock (Tg C) of the reservoir area in a given year.

3. Results

3.1. Interannual Variation in NPP for Different Forest Types

The annual average values of NPP for various forest types in the reservoir area from 1981 to 2014 are shown in Figure 3. The simulation results indicated that the interannual variation of NPP was drastic, presenting five consecutive and alternatively fluctuating curves according to the different forest types. The higher and lower values exhibited consistent variation trends. Coniferous forests accounted for the largest area in the reservoir region and were mainly composed of Pinus massoniana forest, cedar forest, and Chinese fir forest, accounting for 35%-40% of the total forest area. The average NPP value was 550.29 g C m⁻² year⁻¹, ranging between 262.93 and 807.71 g C m⁻² year⁻¹, with a value difference of 544.78 g C m⁻² year⁻¹ between the maximum and the minimum values. Mixed coniferous and broadleaf forests were mainly dominated by Pinus massoniana, cedar, Chinese fir, Quercus variabilis, and oriental white oak, accounting for 15%–18% of the total forest area. The average NPP value was 504.49 g C m⁻² year⁻¹, varying between 307.02 and 654.63 g C m⁻² year⁻¹. The deciduous broadleaf forest in the reservoir region is severely disturbed by human activities and presented a stripped distribution pattern. It accounted for 11% to 14% of the total forest area. The average NPP value was 458.69 g C m⁻² year⁻¹, ranging between 351.12 and 531.62 g C m⁻² year⁻¹. The evergreen broadleaf forest represented the zonal vegetation in the reservoir area, but excessive exploitation had led to serious disturbance and destruction of the original vegetation. The area with natural forest was extremely limited and mainly consists of secondary forests, accounting for only about 5% of the total forest area. The average NPP value was 456.40 g C m⁻² year⁻¹, ranging between 248.69 and 607.67 g C m⁻² year⁻¹. The shrub forest is dominated by Cotinus coggygria, Rhus javanica L., Coriaria nepalensis and Lespedeza bicolor, and its area accounted for about 30% of the total forest area [46]. The average NPP value was 274.99 g C m⁻² year⁻¹, ranging between 160.66 and 484.07 g C m⁻² year⁻¹.

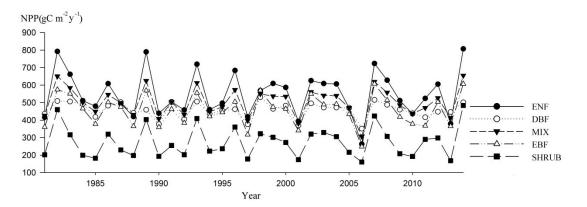


Figure 3. Time series of NPP for various forests in TGRA from 1981 to 2014.

3.2. Temporal Evolution of Forest Productivity and Carbon Sequestration in the TGRA

3.2.1. Forest Productivity

The total forest NPP and the unit area NPP as well as the variation situations of NPP for various forest types during 1992–2012 are shown in Table 5. Both the total and unit area forest NPP in the TGRA presented an obvious interannual fluctuation. In 1992, the values were 7.827 Tg C and 3.833 Mg C hm⁻² year⁻¹, respectively, while in 1996, total and unit area forest NPP values showed similar variations and increased by 44.8 and 41.8% compared to the values in 1992, reaching 11.338 Tg C and 5.437 Mg C hm⁻² year⁻¹. In 2002, total forest NPP continued to increase, albeit with a declining growth rate, while it increased by 17.7% in 2002 compared with the value in 1996. In contrast, the

unit area NPP value decreased slightly in 2002 by 3.72% compared with that in 1996. In 2006, total and unit area NPP values decreased by 46.3 and 53.9% compared with the values in 2002, reaching the minimum values of 7.171 Tg C and 2.415 Mg C hm⁻² year⁻¹, respectively. In 2012, total forest NPP reached the maximum value of 14.999 Tg C, while the unit area NPP also increased to 4.465 Mg C hm⁻² year⁻¹, which was 109.2 and 84.9% higher, respectively, compared to the values in 2006.

From 1992 to 2012, the NPP contribution ratio of each forest type to the total forest NPP was relatively stable. To be specific, the NPP of the coniferous forest accounted for the largest proportion of the total NPP, reaching 43.9%–49.9%. The NPP of the mixed coniferous and the broadleaf forest accounted for the second largest proportion of the total NPP, namely, 16.4%–20.6%, while the proportions of shrub forest and deciduous broadleaf forest were similar, namely, 15.4%–18.4% and 13%–15.4%, respectively. The contribution of the evergreen broadleaf forest was the lowest and accounted for only 3.2%–4.7% of the total NPP. The NPP variation trends of different forest types were relatively consistent and similar to that of the total forest NPP. The period from 1992 to 1996 showed rapid forest growth, with significantly increasing NPP values for all forest types. From 1996 to 2002, the NPP values for all forest types basically showed an increasing trend. In 2006, the NPP values for all forest types significantly decreased. In 2012, apart from the increase in the NPP of deciduous broadleaf forest by 57.48% compared to that in 2006, NPP values of all forest types increased by more than 100% to 7.067 Tg C.

Year -		NPP (Tg C)							
Tear	ENF	MIX	DBF	EBF	Shrub	Total	(Mg C hm ⁻² year ⁻¹)		
1992	3.433	1.611	1.208	0.369	1.206	7.827	3.833		
1996	5.058	1.971	1.476	0.481	2.352	11.338	5.437		
2002	6.656	2.192	1.55	0.474	2.474	13.346	5.235		
2006	3.207	1.338	1.129	0.228	1.269	7.171	2.415		
2012	7.067	2.924	1.778	0.474	2.756	14.999	4.465		

Table 5. NPP of Forest Ecosystem in Three Gorges Reservoir Area in 1992~2012.

3.2.2. Forest Carbon Sequestration

The variations in the carbon sequestration of the forest vegetation layer, the litter layer, and the soil layer as well as the variations in carbon storage and carbon density of the entire area from 1992 to 2012 are shown in Table 6.

From 1992 to 2012, the carbon stocks of the forest vegetation layer, the litter layer, and the soil layer as well as the entire area have similar variation trends. Specifically speaking, the stocks of each layer and the entire area continuously increased from 1992 to 2012, while the growing rate began to slow down and stabilize after the rapid increase. From 1992 to 1996, the carbon stocks of the forest vegetation layer, the litter layer, and the soil layer, and the whole area increased by 2.04%–2.16% from 54.53 Tg, 6.678 Tg, 139.637 Tg and 200.845 Tg in 1992, respectively, to 55.707 Tg, 6.814 Tg, 142.607 Tg, and 204.128 Tg in 1996. From 1996 to 2002, the growth rate was significantly higher, and the carbon stocks for each layer and the total area increased by 23.63%, 23.53%, 21.17%, and 21.92% to 68.872 Tg, 8.419 Tg, 172.795 Tg, and 250.084 Tg, respectively. From 2002 to 2006, the total forest carbon stock kept growing, although the growth rated decreased to 16.27%–17.46%; the carbon stocks of the three layers and the entire area were 80.899 Tg, 9.884 Tg, 200.916 Tg, and 291.699 Tg. The variation trends remained the same from 2006 to 2012, increased by 11.39%–13.13% to 90.117 T g, 11.016 Tg, 227.288 Tg, and 328.421 Tg, respectively, in 2012.

Year —	Carbon Storage (Tg)							
rear –	Vegetation	Litter	Soil	Total				
1992	54.53	6.678	139.637	200.845				
1996	55.707	6.814	142.607	205.128				
2002	68.872	8.417	172.795	250.084				
2006	80.899	9.884	200.916	291.699				
2012	90.117	11.016	227.288	328.421				

Table 6. Carbon storage of the Forest Ecosystem in TGR area from 1992 to 2012.

3.3. Spatial Evolution of Forest Productivity and Carbon Sequestration in the TGRA

As shown in Figure 4, regions with higher forest NPP values from 1992 to 2012 were mainly concentrated in the four counties in western Hubei in the eastern area of the TGRAT, and forest productivity gradually increased from west to east. The most significant change occurred in the counties Wuxi, Wushan, and Yunyang in the eastern hinterland. From 1992 to 1996 and further to 2002, forest NPP rose. Up to 2012, the forest productivities of Wulong County and Fengdu County, located in the Karst Mountains, also significantly improved from 80.046 GgT and 52.313 Gg T in 1992 to 179.673 Gg T and 118.145 Gg T in 2012, respectively. The spatial variation of forest productivity was consistent with that of the forest and the fragmented forest patches. Therefore, increases in forest area and optimization of the forest landscape pattern can promote the improvement of forest productivity.

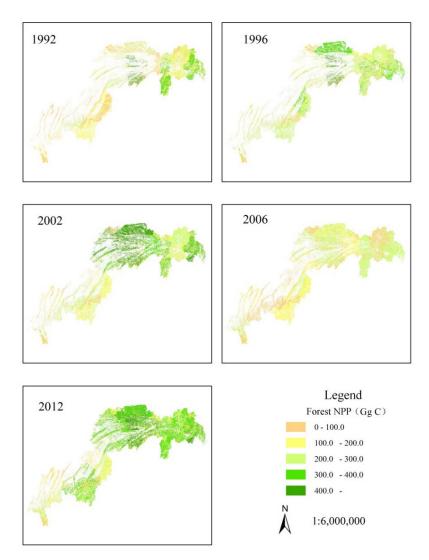


Figure 4. Spatial changes in forest NPP in the TGRA from 1992 to 2012.

As shown in Figure 5, the forest carbon stock in this area also gradually increased from west to east, and the regions with significant carbon stock variation were Wuxi County, Wushan County, and Yunyang County in the eastern hinterland, Wulong County, Fengdu County, and Shizhu County in the western hinterland, and Jiangjin District and Banan District in the western area of the reservoir area. From 1992 to 1996, the forest carbon stock in the eastern hinterland of the reservoir area increased significantly with continuously high values. From 2002 to 2006, the forest carbon stocks of the Jiangjin District and the Banan District, in the hilly areas at the paralleled ridge valley of the western part, increased significantly; while from 1992 to 2012, the forest carbon stocks of Wulong County, Fengdu County, and Shizhu County, in the southwestern hinterland of the reservoir area, increased gradually; this area became the green barrier of the southern reservoir area.

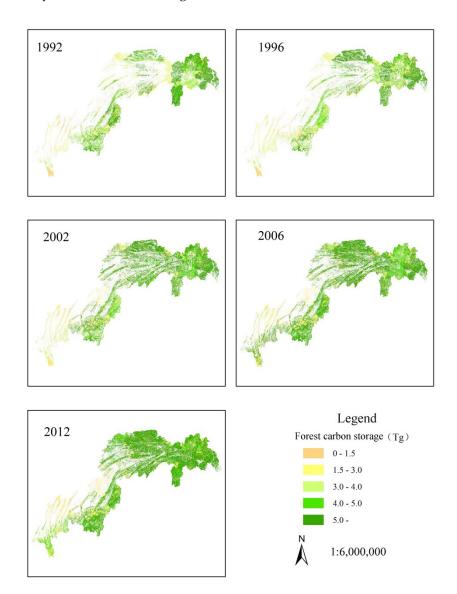


Figure 5. Spatial changes in forest carbon storage in the TGRA from 1992 to 2012.

4. Discussion

4.1. Influences of Air Temperature and Precipitation on Forest Productivity

Forest productivity is not only controlled by the physiological and ecological characteristics of the plants themselves, but is also affected by solar radiation, precipitation, soil, and human activities. From 1981 to 2014, the annual average temperature in the TGRA showed an upward trend in the fluctuating process. During this period, the annual average temperature was 18.49 °C, with slight

annual variations; a minimum temperature of 17.75 °C in 1989 and a maximum temperature of 19.59 °C in 2006. In 2006, the average temperature in China was abnormally high. For most regions within this area, the number of days with a maximum temperature of 35 °C or more was 20 times higher than that in the same period of previous years. In addition, the annual precipitation during 1981 to 2014 in the TGRA showed a larger interannual fluctuation and a slightly downward trend. Average annual precipitation was 1110.34 mm, and the interannual fluctuation was relatively larger, with a maximum annual precipitation of 1415.71 mm (1998) and a minimum annual precipitation of 886.20 mm (2006). Maximum precipitation occurred in 1998, when the Yangtze River Basin suffered a catastrophic flood event; while in 2006, the high temperature and drought event in the Sichuan Province and the Chongqing City led to the decrease of annual precipitation by 259.6 mm compared with that in the same period of "normal" years, and annual precipitation in 2006 was the lowest since 1951 (Figure 6).

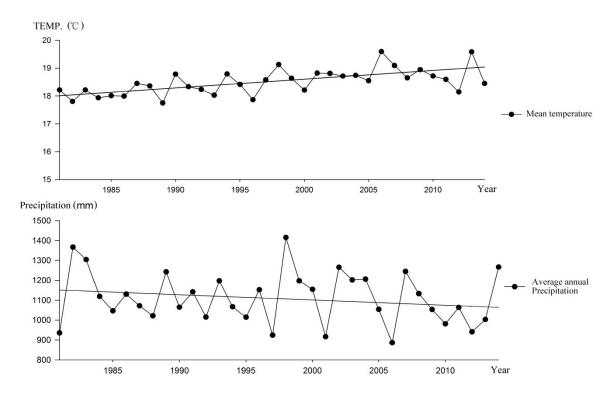


Figure 6. Interannual variation of annual precipitation in Three Gorges Reservoir Area in 1981–2014.

Generally, the air temperature in the TGRA varied slightly, while precipitation presented an increasing trend from 1992 to 2002. With this influence, the unitary forest NPP in the reservoir area firstly increased significantly and then slightly decreased. Pearson's correlation analysis of the NPP values of various forest types and the annual average temperature showed a significantly negative correlation between these factors (Table 7). Besides, Pearson's correlation analysis on the NPP values and the annual precipitation showed a significantly positive correlation (Table 8). As a result of the high temperature and low precipitation in 2006, forest productivity for unit area was extremely low. We therefore assume that unitary forest productivity is greatly affected by the environmental factors such as solar radiation and precipitation in addition to the physiological and ecological characteristics of the plants themselves, whereas total forest productivity is largely influenced by human activities.

	Annual			NPP		
	Temperature	DBF	EBF	ENF	Shrub	MIX
Annual temperature	1	-0.235**	-0.242 **	-0.321 **	-0.248 **	-0.319 **
DBF NPP		1	0.814 **	0.657 **	0.673 **	0.820 **
EBF NPP			1	0.883 **	0.857 **	0.932 **
ENF NPP				1	0.938 **	0.970 **
Sbrub NPP					1	0.929 **
MIX NPP						1

Table 7. Pearson correlation analysis between NPP of forest types and annual mean temperature in TGR.

Note: ** significantly correlated at the 0.01 level (bilateral).

Table 8. Pearson correlation analysis between NPP of five forest types and annual precipitation in TGR.

	Annual			NPP		
	Precipitation	DBF	EBF	ENF	Shrub	MIX
Annual Precipitation	1	0.609 **	0.704 **	0.723 **	0.743 **	0.744 **
DBF NPP		1	0.814 **	0.657 **	0.673 **	0.820 **
EBF NPP			1	0.883 **	0.857 **	0.932 **
ENF NPP				1	0.938 **	0.970 **
Shrub NPP					1	0.929 **
MIX NPP						1

Note: ** significantly correlated at the 0.01 level (bilateral).

4.2. Influences of Land Use Change on Forest Carbon Stocks.

From 1992 to 2012, the areas of various landscape types changed differently [47]. During these 20 years, crop land and forest land were the dominating land use types in this reservoir area, accounting for 93%–96% of the reservoir area. The area of crop land greatly decreased, and its decrease rate was accelerated with the progress of the Three Gorges project. Crop land was mainly transferred to construction land, forest land, water bodies, and the areas of forest land, construction land, or water increased over the years. Among these, construction land showed the most significant increase, increasing five-fold from 1992 to 2012, while water areas in 2012 were two times larger than that in 1992. Forest land increased by nearly 60% or 131.73×10^4 hm² compared to 1992. With the drastic change in land use since the implementation of reservoir project, the continuous increase in total NPP could be attributed to the significant increase in forest area. In addition, the average area of forest patches increased year by year from 1992 to 2012, while patch density decreased. Hereinto, the landscape dominance index of the coniferous forest increased significantly, and the forest landscape pattern changed from small patches to large patches and from a spatially scattered distribution to a continuous distribution, finally enhancing the internal and external connective of the landscape.

Accordingly, from 1992 to 2012, the carbon stocks of the forest vegetation layer, the litter layer, the soil layer, and the entire reservoir area, showed a continuously increasing trend, but the growth rates started to decrease and finally stabilized. The possible reasons could be attributed to the new forest plantation, forest landscape restoration caused by a number of important forestry ecological projects, including the Yangtze River Basin Shelter Forest System project, the Natural Forest Protection Project, and the Grain for Green Project, which were implemented after the severe floods in 1998. From 1992 to 2002, the total forest area in the reservoir region increased from 204.17 × 10^4 hm² to 254.93 × 10^4 hm².

5. Conclusions

The Three Gorges Reservoir area is one of the most ecologically sensitive areas in China, and the forest landscape pattern in this region shows dramatic change due to the influence of the Three Gorges reservoir project. The forest NPP and carbon sequestration of various forest types in this area, which were estimated based on the Biome-BGC model, indicated that total and unitary forest NPP presented obvious annual fluctuation from 1992 to 2012 under the combined influences of land use change and extreme weather events. Apart from the year 2006, from 1992 to 2012, the NPP values of each forest type showed an increasing trend, although the growth rates decreased. In 2006, due to abnormally high air temperatures and less precipitation, total and unit area forest NPP values decreased by 46.3% and 53.9%, respectively, compared to 2002. From 1992 to 2012, the carbon stocks of the forest vegetation layer, the litter layer, the soil layer, and the entire area gradually increased with a decreasing growth rates. Forest carbon stocks were high in the east and the south and low in the west and the north. Regions with higher carbon stocks included the reservoir head, the Wushan-Wuxi section, the Shizhu-Wulong section, and the southern section of the Jiangjin District. From 1992 to 2012, the variation in the forest landscape pattern and the responses of forest productivity and carbon stocks indicated that the increase in forest area and the optimization of the forest landscape pattern could improve the forest productivity and carbon stocks.

In fact, the potential and historic land use changes influence not only the forest areas, but also the forest landscape patterns, such as patch size, fragmentation degree, and patches connectivity and so on. This paper only analyzed the changes in forest productivity and carbon storage caused by the changes of forest area, but did not deeply analyze the impacts of landscapes patterns. Therefore, it is an important issue for further study to clarify the contributions of landscape pattern to carbon storage based on the principle of edge effects.

Based on the Biome-BGC model, the simulated NPP values and carbon density values of the coniferous forest and the mixed coniferous and broadleaf forest were close to the measured values, while those of the deciduous broadleaf forest, the evergreen broadleaf forest, and the shrub forest were lower than the measured values. The possible reason could be the use of the default physiological and ecological parameters provided by the model, as well as the underestimate of the carbon density of forest soil. The effective approach to improve the precision of the Biome-BGC model is to provide more accurate spatial interpolation through the plot investigation data of meteorological stations and more physiological and ecological parameters of different vegetation types (e.g., bamboo forest, forest plantation, etc.). By combination of field plot measurement and model simulation, the conversion from point to region and a multi-scale simulation will be realized; also, the accuracy of scaling up or scaling down will be improved.

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